

Identifying functional flow linkages between stream alteration and biological stream condition indices across California

Ryan Peek^{1*}, Katie Irving², Sarah Yarnell¹, Rob Lusardi^{1, 3}, Eric D. Stein², Raphael Mazor²

¹Center for Watershed Sciences, University of California, Davis, United States, ²Southern California Coastal Water Research Project, United States, ³Department of Wildlife, Fish, and Conservation Biology, University of California, Davis, United States

Submitted to Journal:
Frontiers in Environmental Science

Specialty Section:
Freshwater Science

Article type:
Original Research Article

Manuscript ID:
790667

Received on:
07 Oct 2021

Journal website link:
www.frontiersin.org

Conflict of interest statement

The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest

Author contribution statement

RP led the preparation of this manuscript, data analysis, coding, and project design. KI contributed to data analysis and coding. RP, ES, KI, RM, SY, RL contributed to the conceptualization and project design. All authors contributed to writing this manuscript.

Keywords

bioassessment, flow modification, ecological flow management, Seasonality, Flow-ecology relationship, California Environmental Flows Framework

Abstract

Word count: 279

Large state or regional environmental flow programs, such as the one based on the California Environmental Flows Framework (CEFF), rely on broadly applicable relationships between flow and ecology to inform management decisions. California, despite having high flow and bioassessment data density, has not established relationships between specific elements of the annual hydrograph and biological stream condition. To address this, we spatially and temporally linked USGS gage stations and biological assessment sites in California to identify suitable site pairs for comparisons of streamflow alteration with biological condition at a statewide scale. Flows were assessed using a set of functional flow metrics which provide a comprehensive way to compare alteration and seasonal variation in streamflow across different locations. Biological response was evaluated using the California Stream Condition Index (CSCI) and Algal Stream Condition Index (ASCI), which quantify biological conditions by translating benthic invertebrate or algal resources with watershed-scale environmental data into an overall measure of stream health. These indices provide a consistent statewide standard for interpreting bioassessment data and, thus, a means of quantitatively comparing stream conditions throughout the state. The results indicate that indices of biological stream condition were most closely associated with flow alteration in timing metrics such as fall pulse timing, dry-season timing, and wet season timing. Magnitude metrics such as dry-season baseflow, wet season baseflow, and the fall pulse magnitude were also important drivers of variation, and a metric of seasonality was strongly tied to biological stream conditions, particularly in snowmelt streams. Development of flow criteria under CEFF should consider that alteration to any of these seasonal flow components (e.g., dry-season baseflow, fall pulse flow, wet-season baseflow, spring recession flow) may be important in restructuring biological communities.

Contribution to the field

Large state or regional environmental flow programs rely on broadly applicable relationships between flow and ecology to inform management decisions. California, USA, despite having high flow and bioassessment data density, has not established relationships between specific elements of the annual hydrograph and biological stream condition. We spatially and temporally linked river flow stations and biological assessment sites in California to identify suitable site pairs for comparisons of streamflow alteration with stream health based on biological indices of benthic invertebrates and algae at a statewide and stream class scale. These indices provide a consistent statewide standard for interpreting bioassessment data and, thus, a means of quantitatively comparing stream conditions throughout the state. The results indicate that stream health indices were most closely associated with flow alteration via timing metrics such as seasonality, fall pulse timing, dry-season timing, and wet season timing. Development of ecological flows should consider that alteration to any of these seasonal flow components may be important in restructuring biological communities, and flow management can be implemented using this approach to further identify linkages between flow and biological stream condition.

Funding statement

Funding for this work was provided by the California Wildlife Conservation Board (Agreement WC-434 1849AB). Open access publication fees were provided by the Library at the University of California, 435 Davis, CA.

Ethics statements

Studies involving animal subjects

Generated Statement: No animal studies are presented in this manuscript.

Studies involving human subjects

Generated Statement: No human studies are presented in this manuscript.

Inclusion of identifiable human data

Generated Statement: No potentially identifiable human images or data is presented in this study.

In review

Data availability statement

Generated Statement: Publicly available datasets were analyzed in this study. This data can be found here: Functional Flow Metrics: Natural functional flow metrics for California: rivers.codefornature.org Bioassessment Data: https://www.waterboards.ca.gov/water_issues/programs/swamp/bioassessment/csci_scores_map.html USGS Gage Flow Data: <https://waterdata.usgs.gov/nwis/rt>.

In review

1 Identifying functional flow linkages between stream alteration and 2 biological stream condition indices across California

3 Ryan Peek^{1*}, Katie Irving², Sarah Yarnell¹, Rob Lusardi^{1,3}, Eric D. Stein², Raphael Mazor²

4 ¹Center for Watershed Sciences, University of California, Davis, USA

5 ² Southern California Coastal Water Research Project, California, USA

6 ³ Department of Wildlife, Fish, and Conservation Biology, University of California, Davis, USA

7 * **Correspondence:**

8 Ryan Peek

9 rapeek@ucdavis.edu

10 **Keywords: bioassessment, flow modification, ecological flow management, seasonality, flow-**
11 **ecology relationships, California Environmental Flows Framework (CEFF)**

12 Abstract

13
14 Large state or regional environmental flow programs, such as the one based on the California
15 Environmental Flows Framework (CEFF), rely on broadly applicable relationships between flow and
16 ecology to inform management decisions. California, despite having high flow and bioassessment
17 data density, has not established relationships between specific elements of the annual hydrograph
18 and biological stream condition. To address this, we spatially and temporally linked USGS gage
19 stations and biological assessment sites in California to identify suitable site pairs for comparisons of
20 streamflow alteration with biological condition at a statewide scale. Flows were assessed using a set
21 of functional flow metrics which provide a comprehensive way to compare alteration and seasonal
22 variation in streamflow across different locations. Biological response was evaluated using the
23 California Stream Condition Index (CSCI) and Algal Stream Condition Index (ASCI), which
24 quantify biological conditions by translating benthic invertebrate or algal resources with watershed-
25 scale environmental data into an overall measure of stream health. These indices provide a consistent
26 statewide standard for interpreting bioassessment data and, thus, a means of quantitatively comparing
27 stream conditions throughout the state. The results indicate that indices of biological stream condition
28 were most closely associated with flow alteration in timing metrics such as fall pulse timing, dry-
29 season timing, and wet season timing. Magnitude metrics such as dry-season baseflow, wet season
30 baseflow, and the fall pulse magnitude were also important drivers of variation, and a metric of
31 seasonality was strongly tied to biological stream conditions, particularly in snowmelt streams.
32 Development of flow criteria under CEFF should consider that alteration to any of these seasonal
33 flow components (e.g., dry-season baseflow, fall pulse flow, wet-season baseflow, spring recession
34 flow) may be important in restructuring biological communities.
35

36 1 Introduction

37 Flow alteration is a pervasive and significant issue globally and in California (Poff et al., 2007;
38 Grantham et al., 2014). Over 95% of California's gaged streams have altered flow (Zimmerman et
39 al., 2017), and hydrologic alteration of flow by dams, diversions and urbanization can impact both
40 seasonal and inter-annual flow variability. These flow modifications can impact population
41 connectivity and gene flow, biodiversity, as well as ecological processes (Dudgeon et al., 2006;
42 Yarnell et al., 2010; Carlisle et al., 2011; Peek et al., 2021). While the causes and related impacts of
43 flow alteration are well documented (Poff et al., 2007), significant gaps exist in linking flow
44 management with ecological responses to track current stream conditions, evaluate restoration
45 efficacy, and provide future flow recommendations (Poff and Zimmerman, 2010).

46 A critical component of developing ecological flow needs for management is to identify relationships
47 between specific flow metrics (that represent distinct elements of the annual hydrograph) and
48 measures of biological stream conditions at broad spatial scales. Biological indicators have been
49 widely used for assessing stream conditions, and benthic invertebrates and algae have been
50 successfully used as indicators of stream health in a wide range of studies across the USA (Stevenson
51 and Smol, 2003; Lawrence et al., 2010; Stevenson et al., 2010; Lunde et al., 2013; Stevenson, 2014;
52 Mazor et al., 2016; Steel et al., 2018). Notably, these bioindicators have been extensively used to
53 quantify biological impairment associated with shifts in the environment. For instance, hydrologic
54 alteration or impairment has been shown to strongly influence aquatic benthic invertebrate
55 communities (Poff et al., 2007; Rehn, 2009), and benthic invertebrates have more recently been used
56 to link metrics of hydrologic variability to biological response (Poff and Zimmerman, 2010; Steel et
57 al., 2018). The direct relationship between algae and flow has been reported as limited (Kirkwood et
58 al., 2009; Miller et al., 2009; Schneider et al., 2016), with some exceptions involving algal blooms in
59 large rivers (Cheng et al., 2019; Xia et al., 2020) and directly following a flood (Schneider et al.,
60 2016). However, impacts of flow alteration on water quality (Nilsson and Renöfält, 2008) can also
61 indirectly influence the composition of algal communities (Allan, 2004; Lange et al., 2016), thus
62 these indicators provide a way to assess current conditions in multiple ways.

63 A major challenge to linking specific flow metrics with biological stream conditions is pairing data
64 spatially and temporally and disentangling complex interactions among various flow components.
65 While several monitoring datasets exist across broad spatial and temporal scales, identifying ways to
66 coalesce and synthesize these data in a cogent manner remains difficult. Synthesis of existing
67 monitoring datasets into a single scale or outcome is an important step for providing context to
68 comparing sites, prioritize management actions, and improve monitoring and evaluation of
69 restoration actions. While unified assessment tools have been developed (see Mazor et al., 2016;
70 Beck et al., 2019b; Theroux et al., 2020), it remains difficult to integrate biological and flow metrics
71 across a highly managed and heterogeneous landscape such as California.

72 There are several key datasets that provide data suitable for evaluating links between functional flows
73 and stream health via biological indicators, though each is collected independently, thus sites are not
74 often nested or designed to work in parallel. For biological and biophysical data, the Surface Water
75 Ambient Monitoring Program (SWAMP) is tasked with assessing surface water quality throughout
76 California. The program coordinates water quality monitoring across the state and collects data to
77 support water resource management by the State Water Boards. For example, the data collected by
78 SWAMP's probabilistic Perennial Stream Assessment survey is used to characterize in-stream
79 biological conditions and make estimates about the extent of healthy streams in different regions of
80 the state. These data include several biological indicators, including benthic invertebrates, benthic
81 algae, and measures of physical habitat integrity. In addition, the US Geological Survey (USGS)

Functional flow linkages between stream alteration and stream health

82 National Water Information System (NWIS) is a comprehensive and distributed application that
83 provides a wide range of water data, including daily stream flows from over 28,000 stations across
84 the United States. However, identifying how these datasets may integrate effectively remains an
85 important first step towards a more unified development of ecological flow needs in freshwater
86 systems.

87 California is well suited to test methods for identifying patterns in biological stream conditions
88 associated with flow alteration because of its diversity of climate, geology, hydrology, and land use
89 impacts. To better describe and quantify the different components of a seasonal hydrograph, a
90 functional flow approach provides a standardized hydrologic method to evaluate and compare the
91 role of flows in a stream ecosystem (Yarnell et al., 2020). Twenty-four functional flow metrics
92 (FFM) were developed for California by Yarnell et al (2015, 2020) and comprise five main flow
93 components (fall pulse flow, wet-season baseflow, peak flow, spring recession flow, and dry-season
94 baseflow) of a flow regime, with individual metrics describing the magnitude, timing, frequency,
95 duration and rate of change of each functional component (Appendix Table S1). They are not directly
96 linked to individual/specific organisms/groups, but are associated to specific biological and
97 ecosystem processes (Yarnell et al., 2020). Calculated from existing flow data, functional flow
98 metrics provide a comprehensive way to compare alteration and seasonal variation in streamflow
99 across different locations.

100 Standardized bioassessment indices such as the California Stream Condition Index (CSCI) and the
101 Algal Stream Condition Index (ASCI) are quantitative measures of stream condition which can be
102 used across broad spatial scales (Mazor et al., 2016; Theroux et al., 2020). Leveraging these
103 statewide datasets in conjunction with recent methods for quantifying hydrologic variability at the
104 stream segment scale across California (Stein et al.; Yarnell et al., 2020), provides a unique
105 opportunity to assess biological response to hydrologic alteration in California. ASCI and CSCI are
106 predictive multimetric indices developed for California streams (Mazor et al., 2016; Theroux et al.,
107 2020) and comprise many stream and landscape components that describe biological sensitivities or
108 tolerances to disturbance. The indices allow for the evaluation of biotic response without specificity
109 to one individual metric (e.g., taxa richness), enabling coverage of a broader range of characteristics
110 and stressors associated with individual watersheds. These indices are intended to aid stream
111 management and decision making (e.g., condition assessment, prioritization and flow target
112 development; see (Stein et al., 2017; Mazor et al., 2018; Beck et al., 2019a) and have been integrated
113 into unified assessments of stream health (Beck et al., 2019b). With low regional bias and
114 consideration of natural variation, ASCI and CSCI can distinguish between reference and
115 biologically degraded sites, can be applied at multiple scales, and are appropriate to apply to the
116 diverse landscapes of California (Mazor et al., 2018).

117 Identifying and understanding functional flow linkages between biological responses and specific
118 elements of the flow regime are crucial for implementing a functional flow approach. Identifying
119 which metrics or elements of a flow regime have the greatest impact on benthic invertebrate and algal
120 communities for monitoring and management of stream health provides a quantitative method to
121 build on documented relationships between flow and functional processes in riverine systems
122 (Yarnell et al., 2015). Functional flow metrics provide a method to quantify these linkages—for
123 example the hyporheic zone is linked with wet season baseflow, the fall pulse flow helps flush
124 nutrients downstream, and the spring recession can export nutrients from floodplain to channel—as
125 different hydrologic elements can support different biogeochemical and ecosystem functions (Yarnell
126 et al., 2015).

127 A significant gap in our understanding and implementation of environmental flows is determination
128 of how certain flow components link to biological stream conditions. For assessment, monitoring,
129 and recovery purposes, it is important to determine and evaluate such linkages. Identifying key
130 parameters that are comparable and measurable (i.e., FFM, CSCI, ASCI) is important to assess
131 restoration efficacy and track environmental change in managed freshwater systems, particularly as
132 demand for freshwater increases. Despite this importance, few studies have examined these linkages
133 at broad spatial scales. As such, our objectives were to: (i) identify functional flow-biological
134 condition metrics that explain the greatest variation in statewide and regional data, and (ii) assess
135 relationship trends between functional flow metrics and biological condition. This research has
136 important implications for environmental flow management, particularly where practitioners seek to
137 link biological response to functional flow components.

138 **2 Materials and Methods**

139 **2.1 General Approach**

140 To assess relationships between streamflow condition and stream health, all ASCI and CSCI sites
141 were spatially and temporally paired with proximal USGS gages across California. Using these
142 paired sites, we calculated functional flow metrics for the 24 metrics defined by Yarnell et al. (2020),
143 using a minimum of 10 years of continuous flow data at each selected USGS gage site. In some
144 cases, ASCI and CSCI sites were associated with more than one USGS gage. We calculated a metric
145 of hydrologic alteration using a normalized difference between the observed median value and
146 predicted median value of each metric. Statistical models were then developed to identify which of
147 the functional flow metrics were most closely associated with biological index scores, and the
148 directionality of those relationships.

149 **2.2 Pairing of Biological Stream Condition (CSCI & ASCI) sites with USGS gage sites**

150 We identified all bioassessment sites (n=2,935) in the SWAMP dataset with available ASCI and
151 CSCI scores from data sampled between 1994-2018 during late spring and summer months (May to
152 September, when sampling typically occurs). To pair bioassessment sites with USGS gage sites, we
153 filtered locations to include only bioassessment sites occurring in the same HUC12 catchment as
154 USGS gages with at least 10 years of contiguous daily flow data (Figure 1). We filtered
155 bioassessment sites from the previous step to include only sites on the same National Hydrography
156 Dataset (NHD) mainstem stream or river as the USGS gage (in the same HUC12 watershed)—
157 provided each site was within 10 km downstream of the gage—using the *nhdplusTools*, *dplyr*, and *sf*
158 packages in R version 4.1.1 (Blodgett, 2018; Pebesma, 2018; Wickham et al., 2018, 2019; R Core
159 Team, 2021). Using this list of biological-gage site pairs, we removed sites that did not contain flow
160 data after 1994 to ensure temporal overlap with the biological assessment sampling events (i.e., all
161 ASCI and CSCI data was collected and calculated after 1994). Data from final site pairs were used in
162 all subsequent analyses. For BMI sampling events and resulting ASCI or CSCI scores that occurred
163 in the same water year at the same location, we calculated the median value of these replicate scores
164 to use in the statistical modeling.

165 **2.3 Calculating Delta Hydrology using Functional Flow Metrics**

166 Once the selected ASCI and CSCI sites were paired with proximal USGS sites, we calculated
167 functional flow metrics (Grantham et al.) over the longest contiguous period of record for each USGS

Functional flow linkages between stream alteration and stream health

168 gage using the using the Functional Flows Calculator API client package in R (version 0.9.7.2)¹,
169 which uses hydrologic feature detection algorithms developed by Patterson et al. (2020) and the
170 Python functional flows calculator². We calculated a normalized metric based on the departure from
171 the predicted reference flow (difference between the observed functional flow metric and the
172 predicted functional flow metric) associated with the stream segment at the USGS gage (see
173 Grantham et al. this issue for additional details on how predicted reference-based functional flow
174 metrics were modeled). This measure of delta hydrology was calculated as:

$$\begin{aligned} 175 & \quad (50 \text{ percentile Observed FFM} \\ 176 & \quad \quad - 50 \text{th percentile Predicted FFM}) / 50 \text{th percentile Predicted FFM} \end{aligned}$$

177 In some cases, the functional flow metric value for a single water year at a gage could not be
178 calculated, resulting in an ‘NA’ value. This could occur for several reasons, such as the data record
179 was incomplete, or the annual hydrograph was extremely different compared with the predicted
180 reference condition. These instances would lack a seasonal flow pattern which the flow calculator
181 needs to derive subsequent metrics (see Grantham et al., this issue). If more than 70% of the annual
182 values for a metric across the period of record at a gage were NA, then the flow alteration for that
183 metric at that gage was not included in the dataset. One additional metric, seasonality, was calculated
184 for each gage using the same period of record, based on Colwell’s metrics which measure the
185 seasonal predictability of environmental phenomena (Colwell, 1974). These metrics are defined in
186 terms of Predictability (P), Constancy (C), and Contingency (M)—where M means temporal
187 variability, or seasonality, and P is the reliable recurrence of seasonal patterns across multiple cycles.
188 Importantly, Colwell’s P is maximized when environmental phenomenon is constant throughout the
189 year, if the seasonal fluctuation is consistent across all years, or a combination of both (Tonkin et al.,
190 2017). Following Tonkin et al. (2017), we used calculated seasonality as Colwell’s M/P, as it can be
191 applied in a wide range of ecological studies (Tonkin et al., 2017; Radecki-Pawlik et al., 2020; Peek
192 et al., 2021), and provides a measure (ranging from 0 to 1, with 1 being highly seasonal) of how the
193 environment varies within a single year, which in this case was based on daily flow values from each
194 gage selected for analysis.

¹ https://github.com/ceff-tech/ffc_api_client

² <https://github.com/NoellePatterson/ffc-readme>

195

196 **2.4 Statistical Analysis of Stream Condition Indices vs. Functional Flow Metrics**

197 To determine which functional flow metrics had the strongest association with streamflow alteration,
198 we modeled estimates of delta hydrology (departure from the predicted reference flow) for each
199 functional flow metric against biological condition scores (i.e., ASCI and CSCI) using boosted
200 regression tree analysis, following methods from Steel et al. (2018).

201 Each model was run with CSCI or ASCI as the response, and the delta hydrology statistic for each
202 FFM as well as seasonality as the covariates. Boosted regression trees, a method from the decision
203 tree family of statistics, are well suited for large and complex ecological datasets; they do not assume
204 normality nor linear relationships between predictor and response variables, they ignore non-
205 informative predictor variables, and they can accept predictors that are numeric, categorical, or
206 binary (Elith et al., 2008; Brown et al., 2012). Boosted regression trees are also unaffected by outliers
207 and effectively handle both missing data and collinearity between predictors (De'ath, 2007; Dormann
208 et al., 2013). Importantly, such methods are becoming more common in ecological analyses and have
209 been shown to outperform many traditional statistical methods such as linear regression, generalized
210 linear models, and generalized additive models (Guisan et al., 2007). Boosted regression tree models
211 were run with grid iteration and tuning across parameters (shrinkage [0.001–0.005], interaction depth
212 [3–5], number of minimum observations in a node [3–10], and bag fraction [0.75–0.8]) in model
213 validation, following guidelines from Elith et al. (2008). To assess the relative influence of each
214 functional flow metric in the model, we used the mean-square error method (Ridgeway, 2015).

215 The most influential functional flow metrics were further examined by plotting the delta hydrology
216 metric values against biological condition scores. To better understand regional patterns and assess
217 relationships across different scales, we also analyzed ASCI and CSCI scores and delta hydrology for
218 FFM across three stream classifications—snowmelt, rain, and mixed (combination of rain, snow, or
219 groundwater)—based on Patterson et al (2020) and (Lane et al., 2017) at a regional scale in
220 California. Thus, each model was run using only sites associated with one of these stream classes.

221 **3 Results**

222 **3.1 Pairing of biological stream condition sites with USGS gage sites**

223 We mapped a total of 2,935 unique locations with CSCI values, 2,320 unique locations with ASCI
224 values, and 736 USGS gage sites (Figure 2-3) across California. Despite a relatively large pool of
225 sites to work with, after filtering and pairing, we identified 233 ASCI and 231 CSCI sites associated
226 with 222 USGS gages across the state. Thus, approximately 10% of the total bioindicator sites exist
227 in close spatial proximity (<10 river kilometers) to USGS gage sites with long-term flow data (>10
228 years). Eight metrics were dropped (Appendix Table S1) from the functional flow metric
229 calculations, thus, for every site pair, data included a single bioindicator score, and 16 flow alteration
230 metric scores, one for each of the remaining functional flow metrics. The functional flow calculator
231 returned a wide range of values that indicate the broad array of regional hydrologic conditions across
232 California, including a small percentage (< 2) of extreme outliers that occurred in the 98th percentile
233 or greater of all data (Figure 4).

234 **3.2 Statistical Analysis for Statewide Site Pair Dataset**

235 Boosted regression tree models with delta hydrology and seasonality metrics explained 46% of the
236 deviance in CSCI data, with a cross-validation correlation of 0.678 ($se = 0.019$) and 31% in ASCI
237 with a cross-validation correlation of 0.552 ($se = 0.041$). Of the 16 functional flow metrics included
238 in the model, eight had relative importance values greater than 5%, and Colwell's seasonality metric
239 was consistently one of the top three variables in all models (Figure 5, Table 1). The two most
240 influential functional flow metrics in the statewide model were fall pulse timing (CSCI=13.6,
241 ASCI=12.3% relative influence) and seasonality (CSCI=15.5%, ASCI=7.6%) (Figure 5). Dry season
242 timing was one of the most important variables in the CSCI model, but it was not influential in the
243 ASCI model (Table 1). Three of the top metrics for CSCI related to timing (fall pulse timing,
244 Coldwell's seasonality, and dry-season timing), while the remaining significant metrics were
245 associated with flow magnitudes (wet-season baseflow and fall pulse magnitude) (Table 1, Figure 5).
246 For ASCI, the top metrics were also primarily associated with timing (fall pulse timing, Colwell's
247 seasonality, wet season timing, and spring timing), while other influential metrics were largely
248 associated with flow magnitude (dry-season baseflow, wet-season baseflow, and fall pulse
249 magnitude). When comparing both ASCI and CSCI cumulatively, the strongest metrics were fall-
250 pulse timing and Colwell's seasonality, followed by dry-season baseflow and wet-season timing.
251 Interestingly, the smallest difference in relative importance occurred in the fall pulse magnitude
252 metric (Figure 5, Table 1).

253 Normalized delta hydrology (departure from reference value) for three of the top functional flow
254 metrics was plotted against the ASCI and CSCI scores, grouped by the degree of stream alteration
255 based on thresholds defined by Mazor et al. (2016) and Theroux et al. (2020). Values that fall below
256 zero indicate flow values that are earlier (timing) or decreased (magnitude) from the expected
257 reference condition (Figure 6). Based on the delta hydrology, fall pulse timing occurred earlier than
258 the expected reference condition across all flow alteration threshold categories—though the lowest
259 values typically corresponded with the most altered category—for both ASCI and CSCI. For
260 magnitude metrics, the pattern was more distinct in the fall pulse magnitude metric for ASCI, which
261 showed all but the “Likely intact” scores were reduced from the expected reference condition, and for
262 CSCI, all the “very likely altered” and “likely altered” categories had distributions that were reduced
263 compared to the expected reference conditions (Figure 6). Interestingly, for Colwell's measure of
264 seasonality, there was a consistent positive trend towards higher CSCI and ASCI scores with more
265 predictable and consistent seasonality (recurring intra-annual patterns of temporal variability, e.g.,
266 summer low flow periods and winter floods occurring each year) (Figure 7).

267 3.3 Statistical Analysis by Stream Class

268 Using the paired sites, we split sites based on stream class (Patterson et al., 2020), which were
269 predominantly in stream segments classified as Rain (Snowmelt: ASCI=37, CSCI=55; Mixed:
270 ASCI=88, CSCI=83; and Rain: ASCI=231, CSCI=226). Note, ASCI and CSCI sites paired with
271 multiple proximal USGS gages, thus sample sizes differ from the total number of unique stations
272 (Figure 2-3). Stream class models of delta hydrology showed seasonality, fall pulse, dry season, and
273 wet season flow components were consistently important in all regional models, while spring
274 recession flow was important primarily in the rain and mixed stream class models (Table 1, Figure
275 8). The only regional model that included a peak flow component was the ASCI-snowmelt model.
276 Here, the 10-year flood magnitude had the highest relative influence score for ASCI, but seasonality
277 and fall pulse metrics were stronger drivers of delta hydrology in CSCI (Table 1). The fall pulse
278 timing was the most dominant metric for CSCI in the rain model and ASCI in the model of mixed
279 stream class sites, while the most important metric in the snowmelt model for CSCI was Colwell's
280 seasonality (Table 1, Figure 9).

281 Several metrics with the highest relative influences in the regional stream class models were further
282 examined by plotting the delta hydrology values for each functional flow metric against the paired
283 bioindicator scores by the thresholds identified in each bioassessment index. Figure 9 shows the
284 highly variable nature of the large dataset, which is to be expected given the inherent wide diversity
285 of climate and topography across California. However, trends in the data indicate potential
286 underlying relationships that should be explored further. These data indicate that as seasonality
287 increases, stream condition (ASCI or CSCI) index also increases, though this pattern is most
288 pronounced in the mixed and snowmelt stream classes (Figure 9A). For fall pulse timing, the rain
289 stream class had the greatest number of sites and the clearest pattern, indicating all sites had earlier
290 than the expected reference condition across all flow alteration categories—though the lowest values
291 corresponded with the most altered category—for both ASCI and CSCI. In the mixed stream class
292 and snowmelt stream class, this pattern was less prevalent.

293 **4 Discussion**

294 Linking flow and bioassessment data sheds light on which relationships are important to consider
295 when establishing flow criteria and our results indicate that a functional flows approach is well suited
296 to improve streamflow management in California (Tonkin et al., 2017, 2021). Alteration to seasonal
297 flow components (e.g., spring recession or fall pulse flow) are closely related to stream health and
298 may be important in restructuring biological communities. Specifically, metrics associated with flow
299 timing (including seasonality) were the most influential in linking functional flow metrics to stream
300 condition. Interestingly, while seasonality was the dominant predictor of delta hydrology for CSCI, it
301 was the fourth most informative metric in ASCI, indicating at large spatial scales (e.g., California),
302 there may be differences in sensitivity to seasonal flow changes between invertebrate and algal
303 communities, and that CSCI based on benthic macroinvertebrates may be a more informative index
304 in determining flow alteration-stream condition linkages.

305 **4.1 Timing metrics had the strongest link to biological condition**

306 In both statewide and regional models, timing metrics were the most important, often comprising
307 three or more of the top five metrics. Of the timing metrics, fall pulse timing was the most influential
308 in describing biological changes in the statewide CSCI and ASCI models. Fall pulse timing is
309 strongly correlated with the first fall precipitation event following the dry season baseflow period,
310 occurring between 1st October and 15th December (Patterson et al., 2020). Typically, fall pulse flows
311 begin in November (Ahearn et al., 2004), but may vary widely (Patterson et al., 2020). Fall flushing
312 flows and the timing of such are important in determining the biological condition of streams. For
313 instance, the fall pulse flow is known to provide organic matter and nutrients subsidies to streams
314 from adjacent riparian habitats and, thus, enhance food resources and detrital carbon for foraging
315 invertebrates (Ahearn et al., 2004; Blanckaert et al., 2013). Fall pulse flows are also known to
316 increase invertebrate habitat availability and heterogeneity (Blanckaert et al., 2013; Naman et al.,
317 2016) and reconnect invertebrate communities and population gene flow through dispersal
318 (Townsend and Hildrew, 1976; Mackay, 1992), providing a vital food resource for resident fishes
319 and other higher order consumers. Thus, fall pulse flow timing may be a key factor in re-establishing
320 food web and community connectivity (Elliott, 1973; Nislow et al., 1998; Romaniszyn et al., 2007).
321 During the dry season low flow period, it is common for invertebrates to use the hyporheic zone as a
322 refuge from potentially unsuitable environmental conditions (e.g., temperature) (Wood et al., 2010;
323 Stubbington, 2012) For instance, fall pulse flows are known to reconnect streams with their
324 hyporheic zone and, as a result, decrease water temperature (Yarnell et al., 2020). Invertebrates also
325 use changes in water temperature (generally associated with season shifts) as a cue for fall or winter

Functional flow linkages between stream alteration and stream health

326 emergence (Ward and Stanford, 1982) the timing of which may help synchronize life history events
327 or behavioral adaptations, ultimately increasing reproductive success (Lytle, 2001; Lytle and Poff,
328 2004).

329 Similar to CSCI, fall pulse timing also explained ASCI variability in the statewide model, much of
330 which was driven by mixed rain and snowmelt streams. Fall pulse timing influence was less
331 influential in snowmelt dominated streams, yet still notable. During the dry season low flow period,
332 filamentous algal mats typically become more prevalent and are associated with increases in stream
333 temperature, reduced streamflow velocity, and nutrient enrichment (McIntire, 1966; Poff et al., 1990;
334 Suren et al., 2003). Changes in stream velocity associated with the arrival of the fall pulse flow may
335 scour and effectively remove algal mats, while improving habitat for different algal assemblages or
336 species ultimately flushing organic material downstream. In mixed rain and snow dominated streams,
337 although fall pulse timing was mostly early, a clear improvement of ASCI condition was shown as
338 values approached reference timing (Figure 9B).

339 Dry season baseflow was an important metric for ASCI in rain dominated streams, showing flows
340 both above and below reference condition impacting algal condition (Table 1, Figure 6), which
341 agrees with (Irving et al.) (this issue). There is a greater distribution of flows below the reference
342 condition in the very likely altered class of ASCI versus the other classes. Dry season baseflow is the
343 low flow period of the water year, which begins after the spring recession has stabilized (Patterson et
344 al., 2020). These low flows support algal growth and primary producers by maintaining water
345 temperature and dissolved oxygen (Appendix Table S1, (Yarnell et al., 2020). Low flows can
346 increase algal biomass and cover (Biggs, 1985; Biggs et al., 2005; Schneider and Petrin, 2017), and
347 due to lower velocities and water temperature, algal communities change from a diatom dominated
348 assemblage to a filamentous algae dominated system (Dewson et al., 2007).

349 **4.2 Seasonality and Climate Change**

350 Cumulatively, timing was the dominant component linking biological stream condition with flow,
351 which is an important factor for invertebrates that have evolved in river systems with consistent
352 hydrologic seasonality and predictability. Timing metrics such as wet season timing, dry season
353 timing, spring timing, and seasonality were all influential in the statewide and regional models. Wet
354 season timing relates to the time of the water year when flows are consistently elevated from dry
355 season low flows driven by rain or snow melt (Patterson et al., 2020), while dry-season timing
356 denotes the time of the water year when flows consistently reach baseflow, following the spring
357 recession (Patterson et al 2020). While California's Mediterranean climate integrates a significant
358 amount of interannual variation (Persad et al., 2020), flow regulation has altered patterns of
359 hydrologic seasonality and predictability in many watersheds (Kupferberg et al., 2012; Peek et al.,
360 2021). These patterns are exacerbated by climate change, which research indicates earlier peak flow
361 and snowmelt timing will continue to occur (Kapnick and Hall, 2010), as well as increased volatility
362 and decreased seasonal predictability via more extreme wet and dry events and swings between these
363 extremes (Swain et al., 2018; Persad et al., 2020). Therefore, environmental flow recommendations
364 should consider the ecological flow needs of these communities at scales appropriate to future
365 conditions. Stream health and biological conditions may not improve if existing communities are
366 mismatched to current environmental conditions (Botero et al., 2015). Thus, efforts should focus on
367 river reaches where flow management may provide opportunities to more closely mimic local
368 reference conditions or consider which functional flow component will be the highest priority given
369 management goals.

370 There are many potential factors that cannot be accounted for within modeling frameworks focused
371 solely on the impacts of flow modification. Interactions with stream temperature, ecological
372 dynamics associated with population density and predation, as well as water chemistry and nutrient
373 loads can all play important roles in influencing biological stream condition (Nilsson and Renöfält,
374 2008; Miller et al., 2009; Lange et al., 2016; Schneider et al., 2016). However, the benefit of linking
375 biological indices like CSCI or ASCI with flow is the ability to quantify and assess stream conditions
376 across broad spatial areas, often with very different underlying geography, geology, and watersheds.
377 These indices are designed to be regionally stable and are standardized so they can be compared
378 across large spatial scales (Mazor et al., 2016). This also means it is important to use caution when
379 interpreting regional models for ecological meaning because CSCI and ASCI produce locally
380 relevant reference expectations. For example, landscape heterogeneity and local seasonality could be
381 a strong driver of variation in the data comparisons of models from the same stream class because
382 these sites may occur in very different geographic regions of California. Future modeling approaches
383 at finer scales may benefit from more specific models that can account for important local variables
384 or use individual functional feeding groups or taxa. Nonetheless, identifying key functional flow
385 metrics that can be evaluated more deeply regarding potential thresholds or discrete trends in
386 alteration will help inform the development of ecological flow criteria. This approach can be used to
387 narrow down and identify specific flow metrics that may be most relevant for management by
388 distilling disparate datasets into more useful and discrete information that can be used to aid decision
389 making.

390 **5 Conclusion**

391 Future analyses may leverage this information and approach to focus on more discrete flow-stream
392 condition linkages, with particular attention to temporal lags associated with drought impacts or the
393 sensitivity of biological metrics. More specific hydrologically sensitive biological metrics (e.g., more
394 distinct functional feeding groups in benthic invertebrate data, hydrologically sensitive taxonomic
395 groups, etc.) may provide additional detail for assessment of the impacts of flow alteration on a given
396 stream reach. Furthermore, this approach provides a method to assess these metrics through time, so
397 adaptive approaches to flow management can be implemented, monitored, and revised based on
398 important linkages between flow and biological stream condition.

399 This analysis highlights that despite the information-rich spatial datasets that span much of
400 California, there remains a significant gap in leveraging and layering these datasets in an effective
401 manner. Pairing biological and flow sites spatiotemporally was challenging, and sites were limited
402 across all stream classes, but particularly in snowmelt dominated systems. When data from biological
403 or hydrological time series are limited, alternative approaches can be implemented using modeled
404 streamflow or modeled stream condition indices to predict whether or not flow alteration deviates
405 from reference expectations (Irving et al.; Stein et al., 2017; Mazor et al., 2018; Maloney et al.,
406 2021). Furthermore, ongoing monitoring may benefit from more discrete and targeted sampling to
407 link biological data more accurately with surface flow data. Nonetheless, this current approach
408 provides a novel integration of disparate spatiotemporal datasets and indicates broad relationships
409 can be identified between functional flow metrics and indices of biological stream condition.

410 **6 Conflict of Interest**

411 *The authors declare that the research was conducted in the absence of any commercial or financial*
412 *relationships that could be construed as a potential conflict of interest.*

413 **7 Author Contributions**

414 RP led the preparation of this manuscript, data analysis, coding, and project design. KI contributed to
415 data analysis and coding. RP, ES, KI, RM, SY, RL contributed to the conceptualization and project
416 design. All authors contributed to writing this manuscript.

417 **8 Contribution to the Field**

418 Large state or regional environmental flow programs rely on broadly applicable relationships
419 between flow and ecology to inform management decisions. California, USA, despite having high
420 flow and bioassessment data density, has not established relationships between specific elements of
421 the annual hydrograph and biological stream condition. We spatially and temporally linked river flow
422 stations and biological assessment sites in California to identify suitable site pairs for comparisons of
423 streamflow alteration with stream health based on biological indices of benthic invertebrates and
424 algae at a statewide and stream class scale. These indices provide a consistent statewide standard for
425 interpreting bioassessment data and, thus, a means of quantitatively comparing stream conditions
426 throughout the state. The results indicate that stream health indices were most closely associated with
427 flow alteration via timing metrics such as seasonality, fall pulse timing, dry-season timing, and wet
428 season timing. Development of ecological flows should consider that alteration to any of these
429 seasonal flow components may be important in restructuring biological communities, and flow
430 management can be implemented using this approach to further identify linkages between flow and
431 biological stream condition.

432 **9 Funding**

433 Funding for this work was provided by the California Wildlife Conservation Board (Agreement WC-
434 1849AB). Open access publication fees were provided by the Library at the University of California,
435 Davis, CA.

436 **10 Acknowledgments**

437 We would like to thank members of the California Environmental Flows Framework technical team
438 for providing input and review of this manuscript.

439

440 **11 References**

441

- 442 Ahearn, D. S., Sheibley, R. W., Dahlgren, R. A., and Keller, K. E. (2004). Temporal dynamics of
 443 stream water chemistry in the last free-flowing river draining the western Sierra Nevada,
 444 California. *J. Hydrol.* 295, 47–63.
- 445 Allan, J. D. (2004). Landscapes and Riverscapes: The Influence of Land Use on Stream Ecosystems.
 446 *Annu. Rev. Ecol. Evol. Syst.* 35, 257–284.
- 447 Beck, M. W., Mazor, R. D., Johnson, S., Wisenbaker, K., Westfall, J., Ode, P. R., et al. (2019a).
 448 Prioritizing management goals for stream biological integrity within the developed landscape
 449 context. *Freshw. Sci.* 38, 883–898.
- 450 Beck, M. W., Mazor, R. D., Theroux, S., and Schiff, K. C. (2019b). The Stream Quality Index: A
 451 multi-indicator tool for enhancing environmental management. *Environmental and*
 452 *Sustainability Indicators* 1–2, 100004.
- 453 Biggs, B. (1985). Algae, A blooming nuisance in rivers. *Soil & Water* 2, 27–31.
- 454 Biggs, B. J. F., Nikora, V. I., and Snelder, T. H. (2005). Linking scales of flow variability to lotic
 455 ecosystem structure and function. *River Res. Appl.* 21, 283–298.
- 456 Blanckaert, K., Garcia, X.-F., Ricardo, A.-M., Chen, Q., and Pusch, M. T. (2013). The role of
 457 turbulence in the hydraulic environment of benthic invertebrates. *Ecohydrol.* 6, 700–712.
- 458 Blodgett, D. (2018). *nhdplusTools: Tools for Accessing and Working with the NHDPlus*. Reston, VA:
 459 U.S. Geological Survey Available at: <https://code.usgs.gov/water/nhdplusTools>.
- 460 Botero, C. A., Weissing, F. J., Wright, J., and Rubenstein, D. R. (2015). Evolutionary tipping points
 461 in the capacity to adapt to environmental change. *Proc. Natl. Acad. Sci. U. S. A.* 112, 184–
 462 189.
- 463 Brown, L. R., May, J. T., Rehn, A. C., Ode, P. R., Waite, I. R., and Kennen, J. G. (2012). Predicting
 464 biological condition in southern California streams. *Landsc. Urban Plan.* 108, 17–27.
- 465 Carlisle, D. M., Wolock, D. M., and Meador, M. R. (2011). Alteration of streamflow magnitudes and
 466 potential ecological consequences: a multiregional assessment. *Front. Ecol. Environ.* 9, 264–
 467 270.
- 468 Cheng, B., Xia, R., Zhang, Y., Yang, Z., Hu, S., Guo, F., et al. (2019). Characterization and causes
 469 analysis for algae blooms in large river system. *Sustainable Cities and Society* 51, 101707.
- 470 Colwell, R. K. (1974). Predictability, Constancy, and Contingency of Periodic Phenomena. *Ecology*
 471 55, 1148–1153.
- 472 De’ath, G. (2007). Boosted trees for ecological modeling and prediction. *Ecology* 88, 243–251.
- 473 Dewson, Z. S., James, A. B. W., and Death, R. G. (2007). Stream ecosystem functioning under
 474 reduced flow conditions. *Ecol. Appl.* 17, 1797–1808.

Functional flow linkages between stream alteration and stream health

- 475 Dormann, C. F., Elith, J., Bacher, S., Buchmann, C., Carl, G., Carré, G., et al. (2013). Collinearity: a
476 review of methods to deal with it and a simulation study evaluating their performance.
477 *Ecography*. 36, 27–46.
- 478 Dudgeon, D., Arthington, A. H., Gessner, M. O., Kawabata, Z.-I., Knowler, D. J., Lévêque, C., et al.
479 (2006). Freshwater biodiversity: importance, threats, status and conservation challenges. *Biol.*
480 *Rev. Camb. Philos. Soc.* 81, 163–182.
- 481 Elith, J., Leathwick, J. R., and Hastie, T. (2008). A working guide to boosted regression trees. *J.*
482 *Anim. Ecol.* 77, 802–813.
- 483 Elliott, J. M. (1973). The food of brown and rainbow trout (*Salmo trutta* and *S. gairdneri*) in relation
484 to the abundance of drifting invertebrates in a mountain stream. *Oecologia* 12, 329–347.
- 485 Grantham, T. E., Carlisle, D. M., Howard, J., Lane, B., Lusardi, R., Obester, A., et al. Modeling
486 functional flows in California's rivers. *Frontiers in Environmental Science-Freshwater*
487 *Science*.
- 488 Grantham, T. E., Viers, J. H., and Moyle, P. B. (2014). Systematic Screening of Dams for
489 Environmental Flow Assessment and Implementation. *Bioscience* 64, 1006–1018.
- 490 Guisan, A., Zimmermann, N. E., Elith, J., Graham, C. H., Phillips, S., and Peterson, A. T. (2007).
491 What Matters for Predicting the Occurrences of Trees: Techniques, Data, or Species'
492 Characteristics? *Ecol. Monogr.* 77, 615–630.
- 493 Irving, K., Taniguchi-Quan, K. T., Poresky, A., Wildman, R., Aprahamian, A., River, C., et al. A
494 process for applying effective flow ecology analysis to aid management decision making.
495 *Frontiers in Environmental Science-Freshwater Science*.
- 496 Kapnick, S., and Hall, A. (2010). Observed Climate–Snowpack Relationships in California and their
497 Implications for the Future. *J. Clim.* 23, 3446–3456.
- 498 Kirkwood, A. E., Jackson, L. J., and McCAULEY, E. (2009). Are dams hotspots for *Didymosphenia*
499 *geminata* blooms? *Freshw. Biol.* 54, 1856–1863.
- 500 Kupferberg, S. J., Palen, W. J., Lind, A. J., Bobzien, S., Catenazzi, A., Drennan, J., et al. (2012).
501 Effects of flow regimes altered by dams on survival, population declines, and range-wide
502 losses of California river-breeding frogs. *Conserv. Biol.* 26, 513–524.
- 503 Lane, B. A., Dahlke, H. E., Pasternack, G. B., and Sandoval-Solis, S. (2017). Revealing the Diversity
504 of Natural Hydrologic Regimes in California with Relevance for Environmental Flows
505 Applications. *J. Am. Water Resour. Assoc.* 53, 411–430.
- 506 Lange, K., Townsend, C. R., and Matthaei, C. D. (2016). A trait-based framework for stream algal
507 communities. *Ecol. Evol.* 6, 23–36.
- 508 Lawrence, J. E., Lunde, K. B., Mazor, R. D., Bêche, L. A., McElravy, E. P., and Resh, V. H. (2010).
509 Long-term macroinvertebrate responses to climate change: implications for biological
510 assessment in mediterranean-climate streams. *J. North Am. Benthol. Soc.* 29, 1424–1440.

Functional flow linkages between stream alteration and stream health

- 511 Lunde, K. B., Cover, M. R., Mazor, R. D., Sommers, C. A., and Resh, V. H. (2013). Identifying
512 reference conditions and quantifying biological variability within benthic macroinvertebrate
513 communities in perennial and non-perennial northern California streams. *Environ. Manage.*
514 51, 1262–1273.
- 515 Lytle, D. A. (2001). Disturbance regimes and life-history evolution. *Am. Nat.* 157, 525–536.
- 516 Lytle, D. A., and Poff, N. L. (2004). Adaptation to natural flow regimes. *Trends Ecol. Evol.* 19, 94–
517 100.
- 518 Mackay, R. J. (1992). Colonization by lotic macroinvertebrates: A review of processes and patterns.
519 *Can. J. Fish. Aquat. Sci.* 49, 617–628.
- 520 Maloney, K. O., Carlisle, D. M., Buchanan, C., Rapp, J. L., Austin, S. H., Cashman, M. J., et al.
521 (2021). Linking Altered Flow Regimes to Biological Condition: an Example Using Benthic
522 Macroinvertebrates in Small Streams of the Chesapeake Bay Watershed. *Environ. Manage.*
523 67, 1171–1185.
- 524 Mazor, R. D., May, J. T., Sengupta, A., McCune, K. S., Bledsoe, B. P., and Stein, E. D. (2018). Tools
525 for managing hydrologic alteration on a regional scale: Setting targets to protect stream
526 health. *Freshw. Biol.* 63, 786–803.
- 527 Mazor, R. D., Rehn, A. C., Ode, P. R., Engeln, M., Schiff, K. C., Stein, E. D., et al. (2016).
528 Bioassessment in complex environments: designing an index for consistent meaning in
529 different settings. *Freshw. Sci.* 35, 249–271.
- 530 McGill, R., Tukey, J. W., and Larsen, W. A. (1978). Variations of Box Plots. *Am. Stat.* 32, 12–16.
- 531 McIntire, C. D. (1966). Some effects of current velocity on periphyton communities in laboratory
532 streams. *Hydrobiologia* 27, 559–570.
- 533 Miller, M. P., McKnight, D. M., Cullis, J. D., Greene, A., Vietti, K., and Liptzin, D. (2009). Factors
534 controlling streambed coverage of *Didymosphenia geminata* in two regulated streams in the
535 Colorado Front Range. *Hydrobiologia* 630, 207–218.
- 536 Naman, S. M., Rosenfeld, J. S., and Richardson, J. S. (2016). Causes and consequences of
537 invertebrate drift in running waters: from individuals to populations and trophic fluxes. *Can.*
538 *J. Fish. Aquat. Sci.* 73, 1292–1305.
- 539 Nilsson, C., and Renöfält, B. (2008). Linking Flow Regime and Water Quality in Rivers: a Challenge
540 to Adaptive Catchment Management. *Ecol. Soc.* 13. doi:10.5751/ES-02588-130218.
- 541 Nislow, K. H., Folt, C., and Seandel, M. (1998). Food and foraging behavior in relation to
542 microhabitat use and survival of age-0 Atlantic salmon. *Can. J. Fish. Aquat. Sci.* 55, 116–127.
- 543 Patterson, N. K., Lane, B. A., Sandoval-Solis, S., Pasternack, G. B., Yarnell, S. M., and Qiu, Y.
544 (2020). A hydrologic feature detection algorithm to quantify seasonal components of flow
545 regimes. *J. Hydrol.* 585, 124787.

Functional flow linkages between stream alteration and stream health

- 546 Pebesma, E. (2018). Simple Features for R: Standardized Support for Spatial Vector Data. *R J.* 10,
547 439.
- 548 Peek, R. A., O'Rourke, S. M., and Miller, M. R. (2021). Flow modification associated with reduced
549 genetic health of a river-breeding frog, *Rana boylei*. *Ecosphere* 12. doi:10.1002/ecs2.3496.
- 550 Persad, G. G., Swain, D. L., Kouba, C., and Ortiz-Partida, J. P. (2020). Inter-model agreement on
551 projected shifts in California hydroclimate characteristics critical to water management. *Clim.*
552 *Change*. doi:10.1007/s10584-020-02882-4.
- 553 Poff, N. L., Olden, J. D., Merritt, D. M., and Pepin, D. M. (2007). Homogenization of regional river
554 dynamics by dams and global biodiversity implications. *Proc. Natl. Acad. Sci. U. S. A.* 104,
555 5732–5737.
- 556 Poff, N. L., Voelz, N. J., Ward, J. V., and Lee, R. E. (1990). Algal Colonization under Four
557 Experimentally-Controlled Current Regimes in High Mountain Stream. *J. North Am. Benthol.*
558 *Soc.* 9, 303–318.
- 559 Poff, N. L., and Zimmerman, J. K. H. (2010). Ecological responses to altered flow regimes: a
560 literature review to inform the science and management of environmental flows: Review of
561 altered flow regimes. *Freshw. Biol.* 55, 194–205.
- 562 R Core Team (2021). *R: A Language and Environment for Statistical Computing*. Vienna, Austria: R
563 Foundation for Statistical Computing Available at: <https://www.R-project.org/>.
- 564 Radecki-Pawlik, A., Wałęga, A., Młyński, D., Młoczek, W., Kokoszka, R., Tokarczyk, T., et al.
565 (2020). Seasonality of mean flows as a potential tool for the assessment of ecological
566 processes: Mountain rivers, Polish Carpathians. *Sci. Total Environ.* 716, 136988.
- 567 Rehn, A. C. (2009). Benthic macroinvertebrates as indicators of biological condition below
568 hydropower dams on west slope Sierra Nevada streams, California, USA. *River Res. Appl.* 25,
569 208–228.
- 570 Ridgeway, G. (2015). *gbm: Generalized Boosted Regression Models*. [http://CRAN.R-](http://CRAN.R-project.org/package=gbm)
571 [project.org/package=gbm](http://CRAN.R-project.org/package=gbm).
- 572 Romaniszyn, E. D., Hutchens, J. J., and Bruce Wallace, J. (2007). Aquatic and terrestrial invertebrate
573 drift in southern Appalachian Mountain streams: implications for trout food resources.
574 *Freshw. Biol.* 52, 1–11.
- 575 Schneider, S. C., Hilt, S., Vermaat, J. E., and Kelly, M. (2016). “The ‘Forgotten’ Ecology Behind
576 Ecological Status Evaluation: Re-Assessing the Roles of Aquatic Plants and,” in *Progress in*
577 *Botany* (unknown).
- 578 Schneider, S. C., and Petrin, Z. (2017). Effects of flow regime on benthic algae and
579 macroinvertebrates - A comparison between regulated and unregulated rivers. *Sci. Total*
580 *Environ.* 579, 1059–1072.

Functional flow linkages between stream alteration and stream health

- 581 Steel, A. E., Peek, R. A., Lusardi, R. A., and Yarnell, S. M. (2018). Associating metrics of
582 hydrologic variability with benthic macroinvertebrate communities in regulated and
583 unregulated snowmelt-dominated rivers. *Freshw. Biol.* 63, 844–858.
- 584 Stein, E. D., Sengupta, A., Mazor, R. D., McCune, K., Bledsoe, B. P., and Adams, S. (2017).
585 Application of regional flow-ecology relationships to inform watershed management
586 decisions: Application of the ELOHA framework in the San Diego River watershed,
587 California, USA. *Ecohydrol.* 10, e1869.
- 588 Stein, E. D., Zimmerman, J., Yarnell, S., Stanford, B., Lane, B., Taniguchi-Quan, K. T., et al. The
589 California Environmental Flows Framework: Meeting the Challenges of Developing a Large-
590 Scale Environmental Flows Program. *Frontiers in Environmental Science-Freshwater
591 Science*.
- 592 Stevenson, J. (2014). Ecological assessments with algae: a review and synthesis. *J. Phycol.* 50, 437–
593 461.
- 594 Stevenson, R. J., Pan, Y., and van Dam, H. (2010). “Assessing environmental conditions in rivers
595 and streams with diatoms,” in *The Diatoms: Applications to the Environmental and Earth
596 Sciences*, eds. E. F. Stoermer and J. P. Smol (Cambridge: Cambridge University Press), 57–
597 85.
- 598 Stevenson, R. J., and Smol, J. P. (2003). “Use of algae in environmental assessments,” in *Freshwater
599 Algae in North America: Classification and Ecology*, eds. J. D. Wehr and R. G. Sheath (San
600 Diego, California: Academic Press, San Diego, California), 775– 804.
- 601 Stubbington, R. (2012). The hyporheic zone as an invertebrate refuge: A review of variability in
602 space, time, taxa and behaviour. *Mar. Freshwater Res.* 63, 293–311.
- 603 Suren, A. M., Biggs, B. J. F., Kilroy, C., and Bergey, L. (2003). Benthic community dynamics during
604 summer low-flows in two rivers of contrasting enrichment 1. Periphyton. *N. Z. J. Mar.
605 Freshwater Res.* 37, 53–70.
- 606 Swain, D. L., Langenbrunner, B., David Neelin, J., and Hall, A. (2018). Increasing precipitation
607 volatility in twenty-first-century California. *Nat. Clim. Chang.* 8, 427–433.
- 608 Theroux, S., Mazor, R. D., Beck, M. W., Ode, P. R., Stein, E. D., and Sutula, M. (2020). Predictive
609 biological indices for algae populations in diverse stream environments. *Ecol. Indic.* 119,
610 106421.
- 611 Tonkin, J. D., Bogan, M. T., Bonada, N., Rios-Touma, B., and Lytle, D. A. (2017). Seasonality and
612 predictability shape temporal species diversity. *Ecology* 98, 1201–1216.
- 613 Tonkin, J. D., Olden, J. D., Merritt, D. M., Reynolds, L. V., Rogosch, J. S., and Lytle, D. A. (2021).
614 Designing flow regimes to support entire river ecosystems. *Front. Ecol. Environ.* 19, 326–
615 333.
- 616 Townsend, C. R., and Hildrew, A. G. (1976). Field Experiments on the Drifting, Colonization and
617 Continuous Redistribution of Stream Benthos. *J. Anim. Ecol.* 45, 759–772.

Functional flow linkages between stream alteration and stream health

- 618 Ward, J. V., and Stanford, J. A. (1982). Thermal Responses in the Evolutionary Ecology of Aquatic
619 Insects. *Annu. Rev. Entomol.* 27, 97–117.
- 620 Wickham, H., Averick, M., Bryan, J., Chang, W., McGowan, L. D., François, R., et al. (2019).
621 Welcome to the tidyverse. *Journal of Open Source Software* 4, 1686.
622 doi:10.21105/joss.01686.
- 623 Wickham, H., François, R., Henry, L., and Müller, K. (2018). dplyr: A Grammar of Data
624 Manipulation. Available at: <https://CRAN.R-project.org/package=dplyr>.
- 625 Wood, P. J., Boulton, A. J., Little, S., and Stubbington, R. (2010). Is the hyporheic zone a refugium
626 for aquatic macroinvertebrates during severe low flow conditions? *Fundam Appl Limnol*
627 *Arch Hydrobiol. Fundamental and Applied Limnology / Archiv für Hydrobiologie* 176, 377–
628 390.
- 629 Xia, R., Wang, G., Zhang, Y., Yang, P., Yang, Z., Ding, S., et al. (2020). River algal blooms are well
630 predicted by antecedent environmental conditions. *Water Res.* 185, 116221.
- 631 Yarnell, S. M., Petts, G. E., Schmidt, J. C., Whipple, A. A., Beller, E. E., Dahm, C. N., et al. (2015).
632 Functional Flows in Modified Riverscapes: Hydrographs, Habitats and Opportunities.
633 *Bioscience* 65, 963–972.
- 634 Yarnell, S. M., Stein, E. D., Webb, J. A., Grantham, T., Lusardi, R. A., Zimmerman, J., et al. (2020).
635 A functional flows approach to selecting ecologically relevant flow metrics for environmental
636 flow applications. *River Res. Appl.* 36, 318–324.
- 637 Yarnell, S. M., Viers, J. H., and Mount, J. F. (2010). Ecology and Management of the Spring
638 Snowmelt Recession. *Bioscience* 60, 114–127.
- 639 Zimmerman, J. K. H., Carlisle, D. M., May, J. T., Klausmeyer, K. R., Grantham, T. E., Brown, L. R.,
640 et al. (2017). Patterns and magnitude of flow alteration in California, USA. *Freshw. Biol.* 16,
641 1311.
- 642
- 643

Functional flow linkages between stream alteration and stream health

644 **13 Tables**
 645
 646 **Table 1.** Mean relative influence values for functional flow metrics included in all CA and each of
 647 the three stream class-based models that assessed flow alteration in relation to ASCI & CSCI scores.
 648 Metrics and relative influence values in gray boxes and bolded were most influential (> 5%).
 649

Flow Metric Name	All CA		Rain		Mixed		Snowmelt	
	CSCI	ASCI	CSCI	ASCI	CSCI	ASCI	CSCI	ASCI
Fall pulse timing	13.6	12.3	20.2	3.6	2.6	35.7	8.6	10.4
Fall pulse magnitude	6.4	6.9	6.3	4.9	2.7	8.1	10	5.9
Wet-season timing	5.1	13.8	2.8	17.6	3.6	6.5	2.1	11.6
Wet-season baseflow	5.8	5	6.6	4.4	1.1	4.6	1.1	5.8
Wet-season duration	4.4	2.7	4.4	3.4	9.7	8.1	5.1	2.5
Wet-season median flow	2.2	3.7	2.7	5.7	3.2	0.7	8.2	1.9
10-year flood magnitude	3.8	3.1	3.5	2.7	5.3	3.3	3.4	16.7
2-year flood magnitude	4.8	2.8	5.2	2.3	3.6	3.1	4.1	2.5
5-year flood magnitude	3	1.4	3.4	1.1	3.3	1.2	3.6	0.9
Spring timing	4.4	4.1	8	6.8	9.1	1.5	1.7	8.4
Spring duration	3.8	4	3.5	2.2	8	7.3	2.7	2.4
Spring recession magnitude	3.8	6.5	2.8	5.8	3	6.7	7.6	1.5
Dry-season high baseflow	2.7	5	3	5.8	9	1	1.2	9.7
Dry-season baseflow	5.9	15.8	6.5	16.4	4.7	3.2	7.4	2.8
Dry-season timing	9.7	1.2	5.1	1.1	11.4	2.7	5	2.6
Dry-season duration	5.2	4	6.2	3.5	5.7	4	8	1.4
Colwell's M/P	15.5	7.6	9.8	12.7	14.2	2.1	20.3	13

650

651 14 Figures

652

653 Figure 1. Flow diagram of steps used to pair biological stream condition sites and USGS gage
654 locations.

655

656 Figure 2. Map of all sampling sites (A) possible CSCI sites and ASCI sites, (B) the potential gages
657 (n=2097) for the final site pairs, and (C) all unique CSCI, ASCI, and USGS gages sites with >10
658 years of flow data. Note, some ASCI and CSCI sites paired with more than one gage site.

659

660 Figure 3. Map of selected biological sampling sites for unique ASCI (circles) and CSCI (diamond)
661 data overlaying stream classifications adapted from Patterson et al. 2020.

662

663 Figure 4. Boxplot of delta hydrology of functional flow metrics used in the analysis across all gages
664 for ASCI and CSCI. The solid pink line in the background indicates no difference between the
665 observed 50th percentile and the predicted reference 50th percentile metric value. Values to the left
666 of the line are reduced or early, values to the right are inflated or late, relative to the expected
667 reference value. Extreme outliers (>98 percentile) have been removed from the boxplot.

668

669 Figure 5. Relative importance of functional flow metrics in boosted regression tree models assessing
670 flow alteration relative to ASCI and CSCI scores for paired sites statewide. Relative influence values
671 were calculated using a mean-square error (MSE) approach, which determines those variables with
672 the largest average reduction in MSE. Functional flow metrics are described in Appendix Table S1.

673

674 Figure 6. Top FFM (normalized as Delta Hydrology) values vs binned ASCI and CSCI values (based
675 on thresholds from Mazor et al., 2016 and Theroux et al., 2020) for all CA site pairs. The red zero
676 line delineates departure from expected reference flow metric, values < 0 are reduced or early, values
677 > 0 are inflated or late, relative to the expected reference value. Notches indicate an approximate
678 95% confidence interval to compare medians, thus if notches of two boxplots do not overlap this
679 suggests the medians are significantly different (see McGill et al., 1978).

680

681 Figure 7. Colwell's seasonality versus binned ASCI and CSCI values (based on thresholds from
682 Mazor et al., 2016 and Theroux et al., 2020) for all CA site pairs. Notches indicate an approximate
683 95% confidence interval to compare medians, thus if notches of two boxplots do not overlap this
684 suggests the medians are significantly different (see McGill et al., 1978).

685

686 Figure 8. Relative importance of functional flow metrics in boosted regression tree models assessing
687 flow alteration relative to ASCI and CSCI scores by stream classification (Patterson et al., 2020).
688 Relative influence values were calculated using a mean-square error (MSE) approach, which
689 determines those variables with the largest average reduction in MSE. Functional flow metrics are
690 described in Table 1.

691

692 Figure 9A. Colwell's seasonality versus binned ASCI and CSCI values (based on thresholds from
693 Mazor et al., 2016 and Theroux et al., 2020) for all site pairs by stream class. Notches indicate an
694 approximate 95% confidence interval to compare medians, thus if notches of two boxplots do not
695 overlap this suggests the medians are significantly different (see McGill et al., 1978).

696

697 Figure 9B. Fall pulse timing (normalized as Delta Hydrology) values vs binned ASCI and CSCI
698 values (based on thresholds from Mazor et al., 2016 and Theroux et al., 2020) for all site pairs by
699 stream class. The red zero line delineates departure from expected reference flow metric, values < 0
700 are reduced or early, values > 0 are inflated or late, relative to the expected reference value.

In review

Figure 1.TIFF

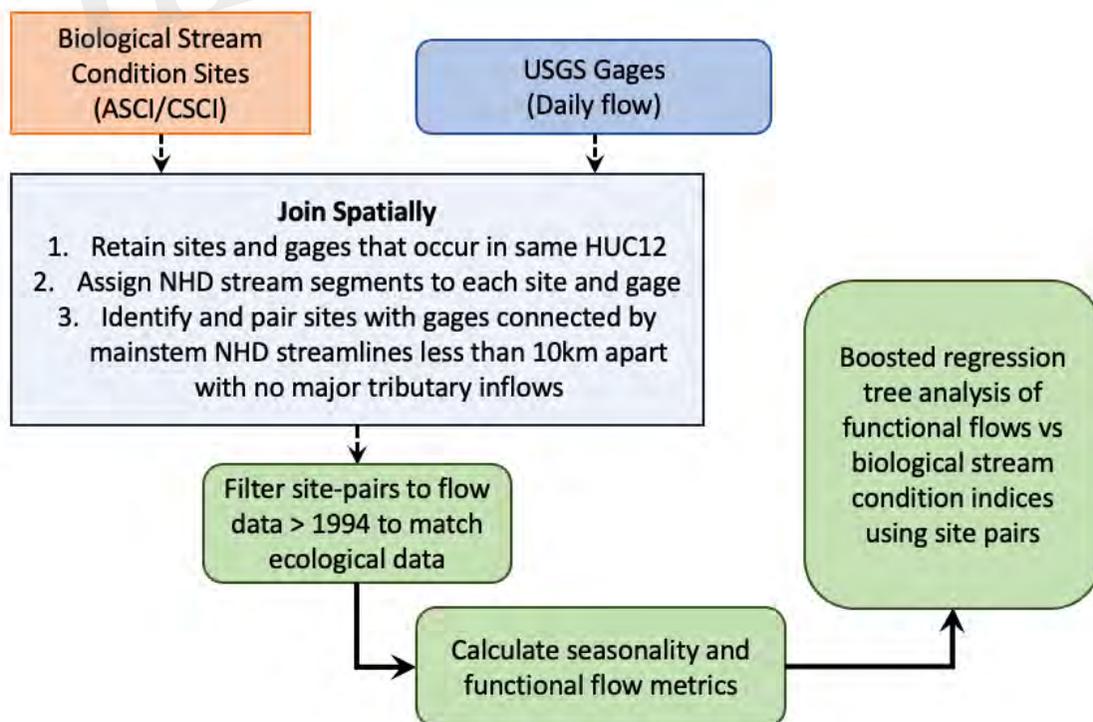


Figure 2.TIFF

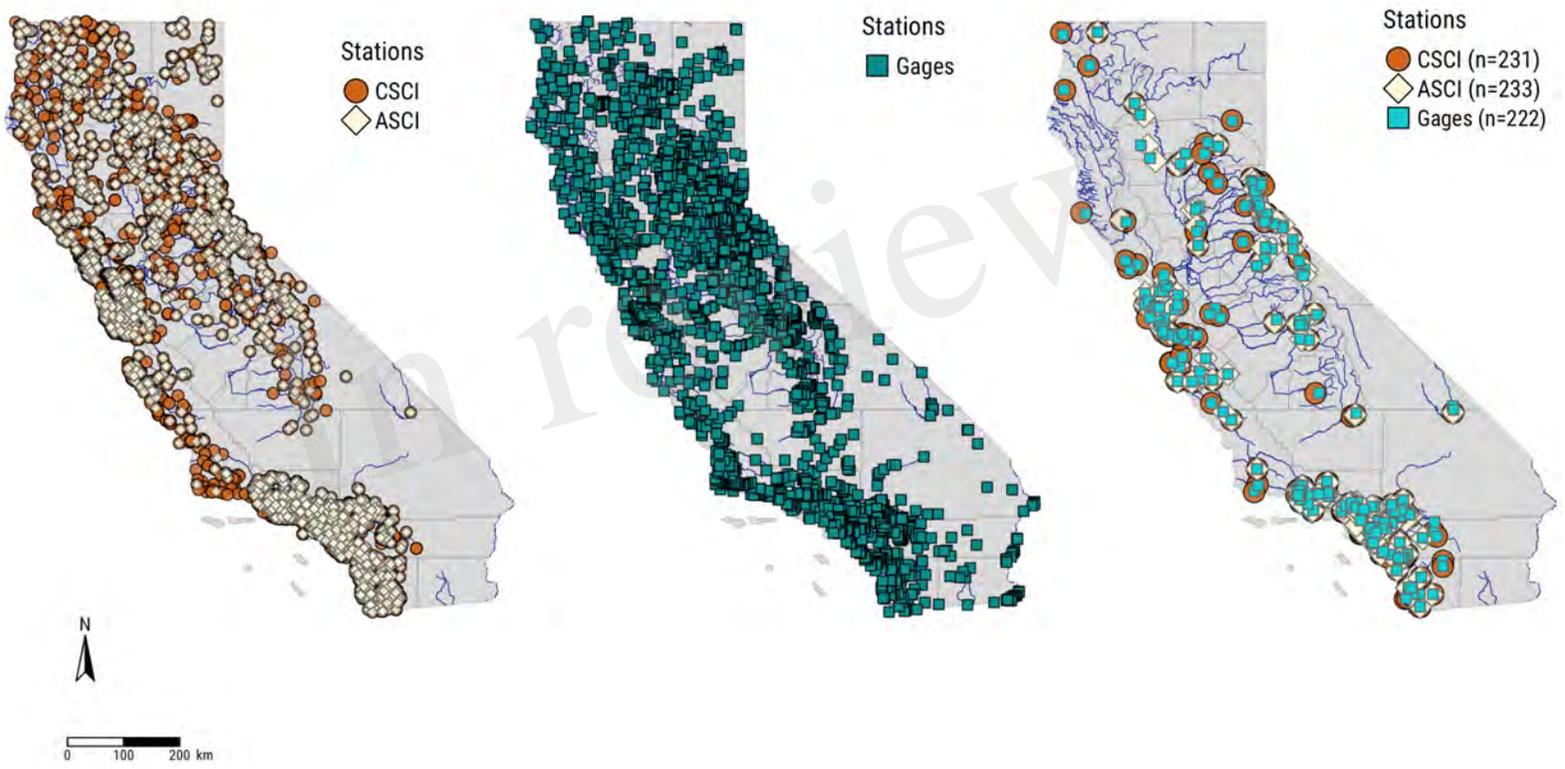


Figure 3.TIFF

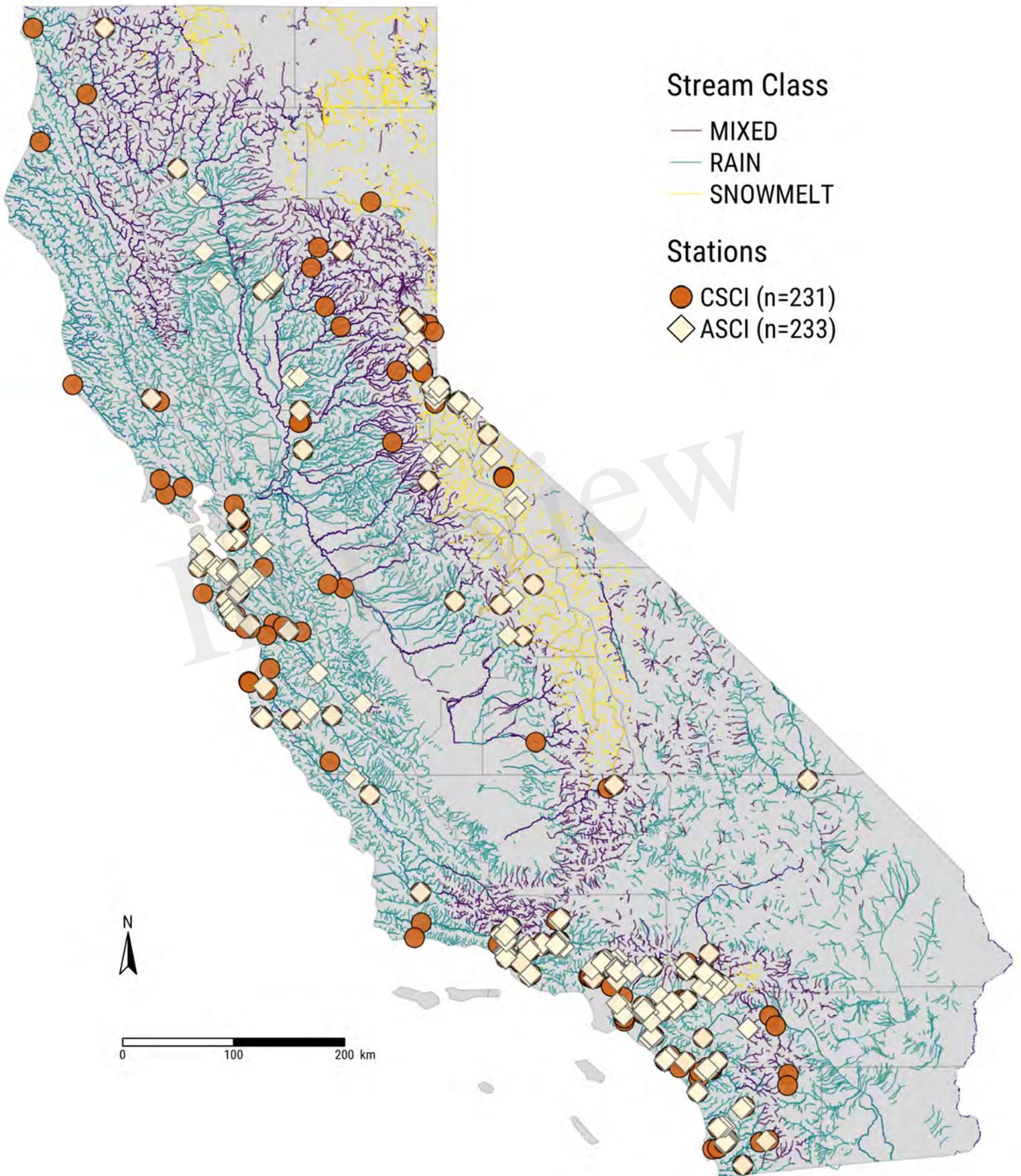


Figure 4.TIFF

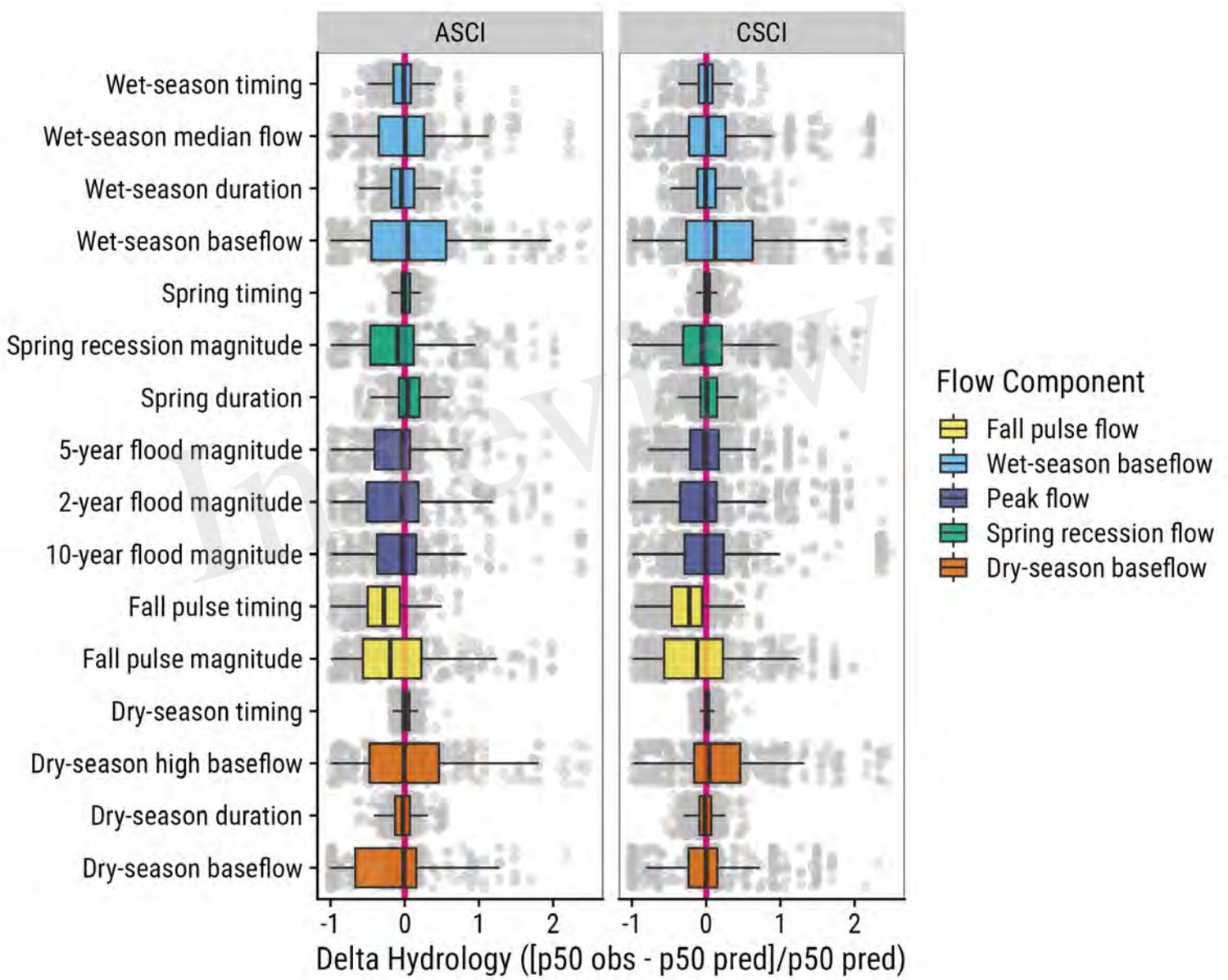


Figure 5.TIFF

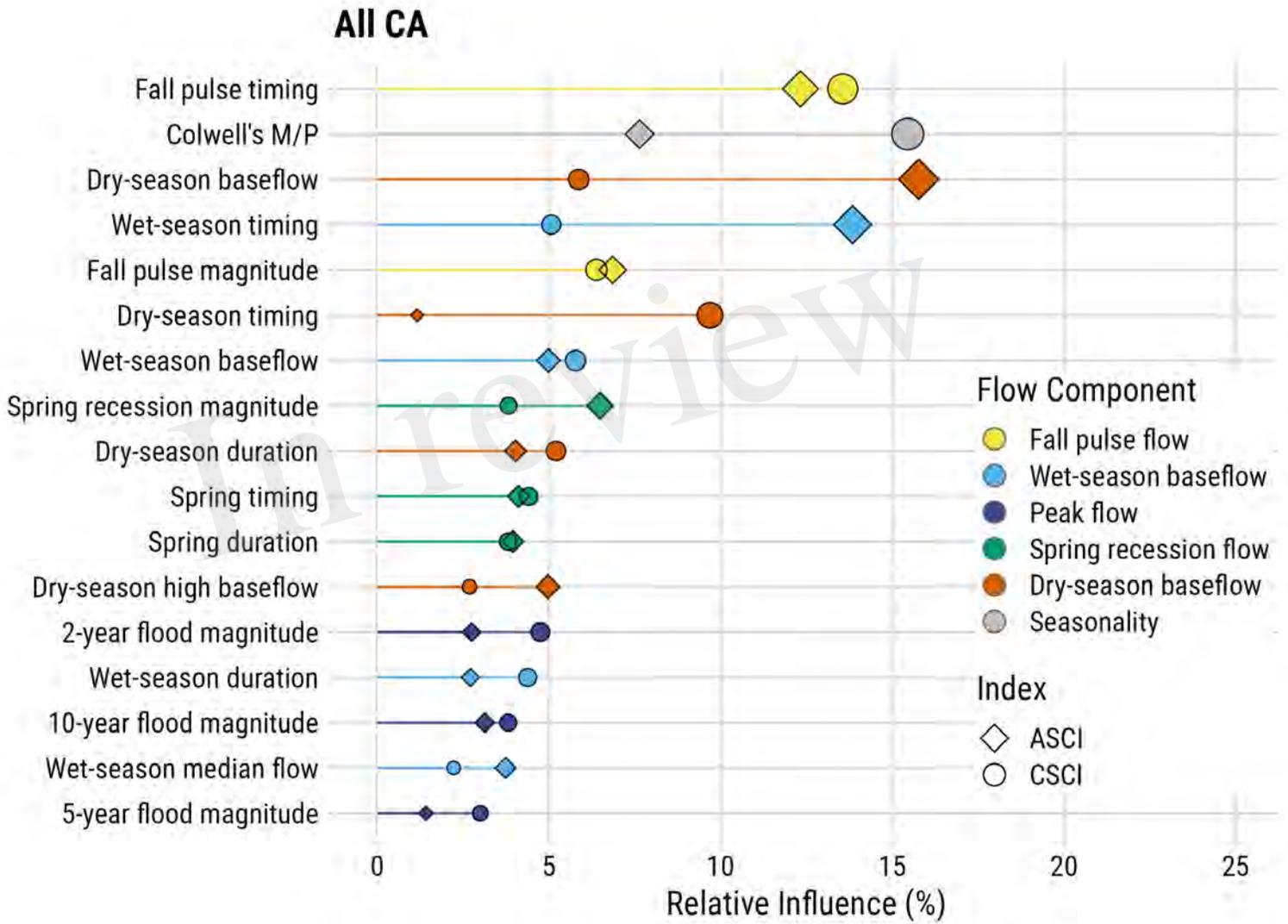


Figure 6.TIFF

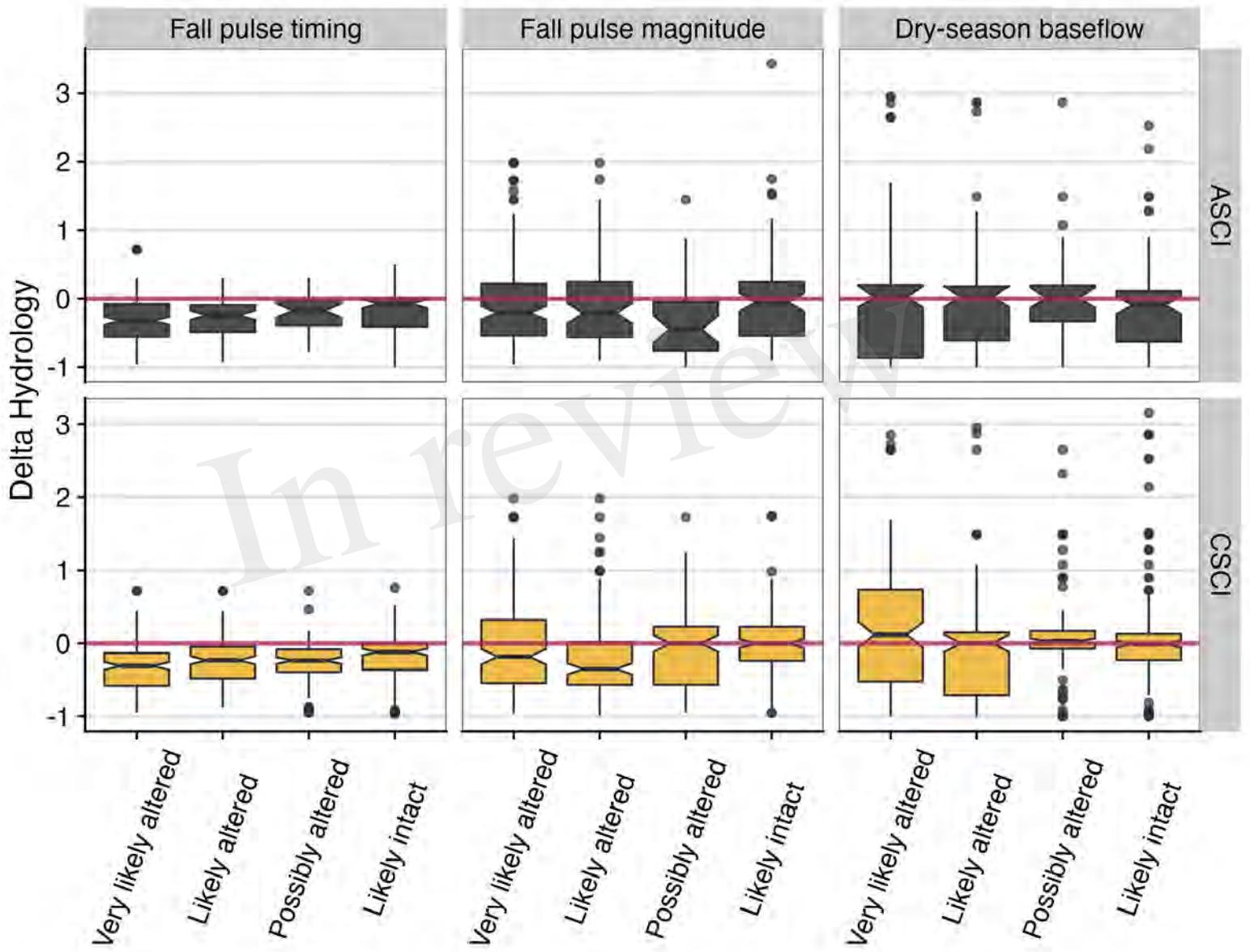


Figure 7.TIFF

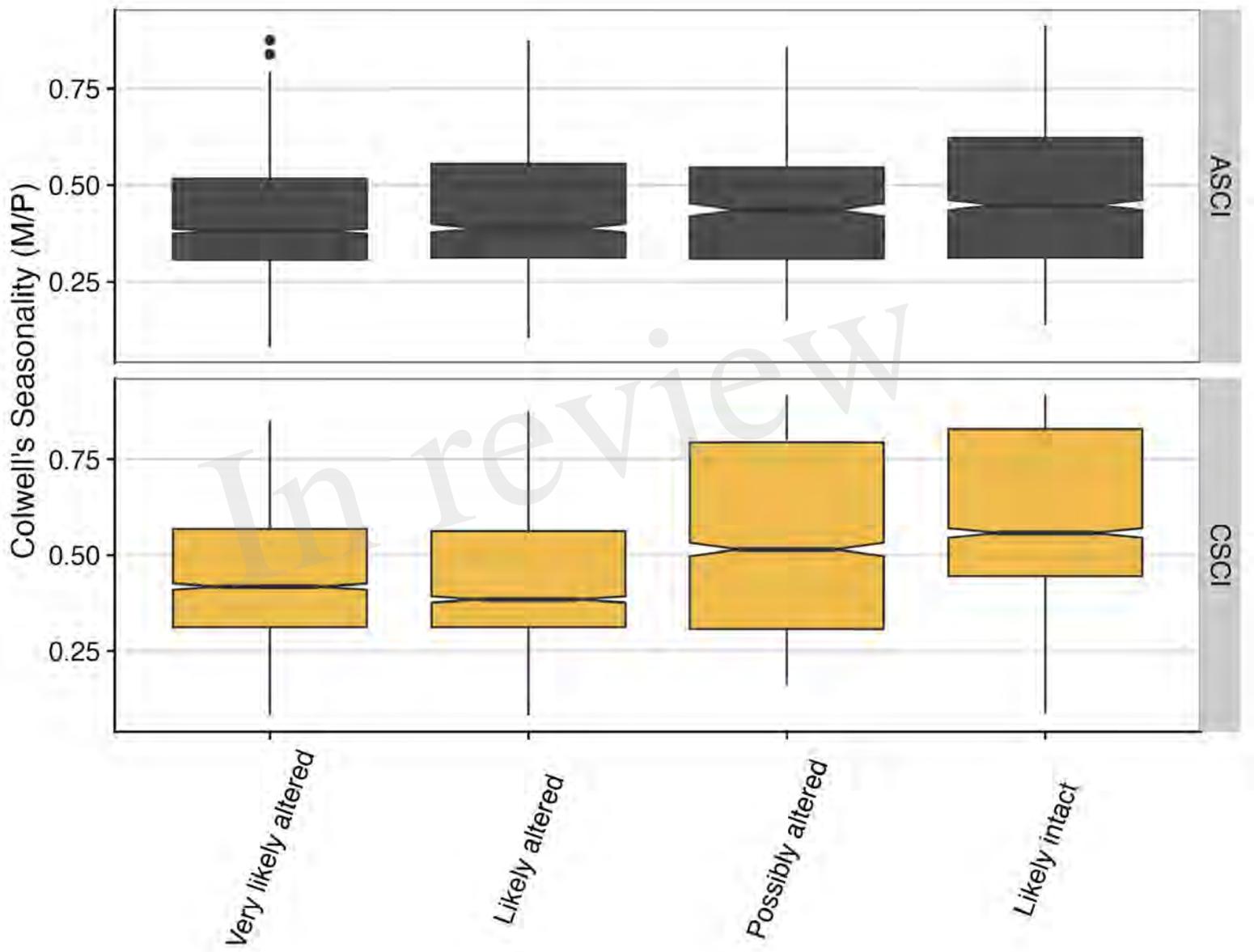


Figure 8.JPEG

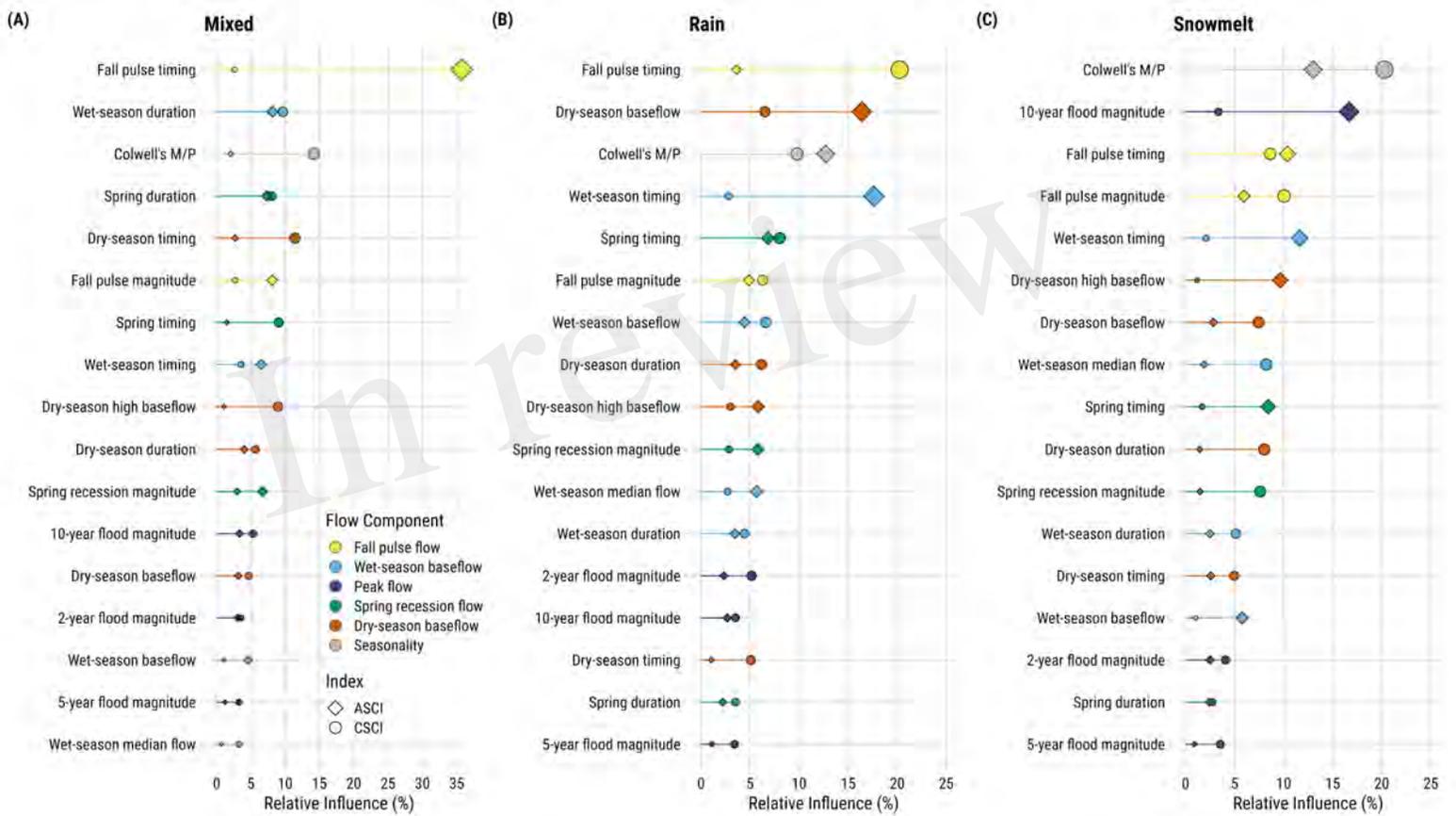


Figure 9.TIFF

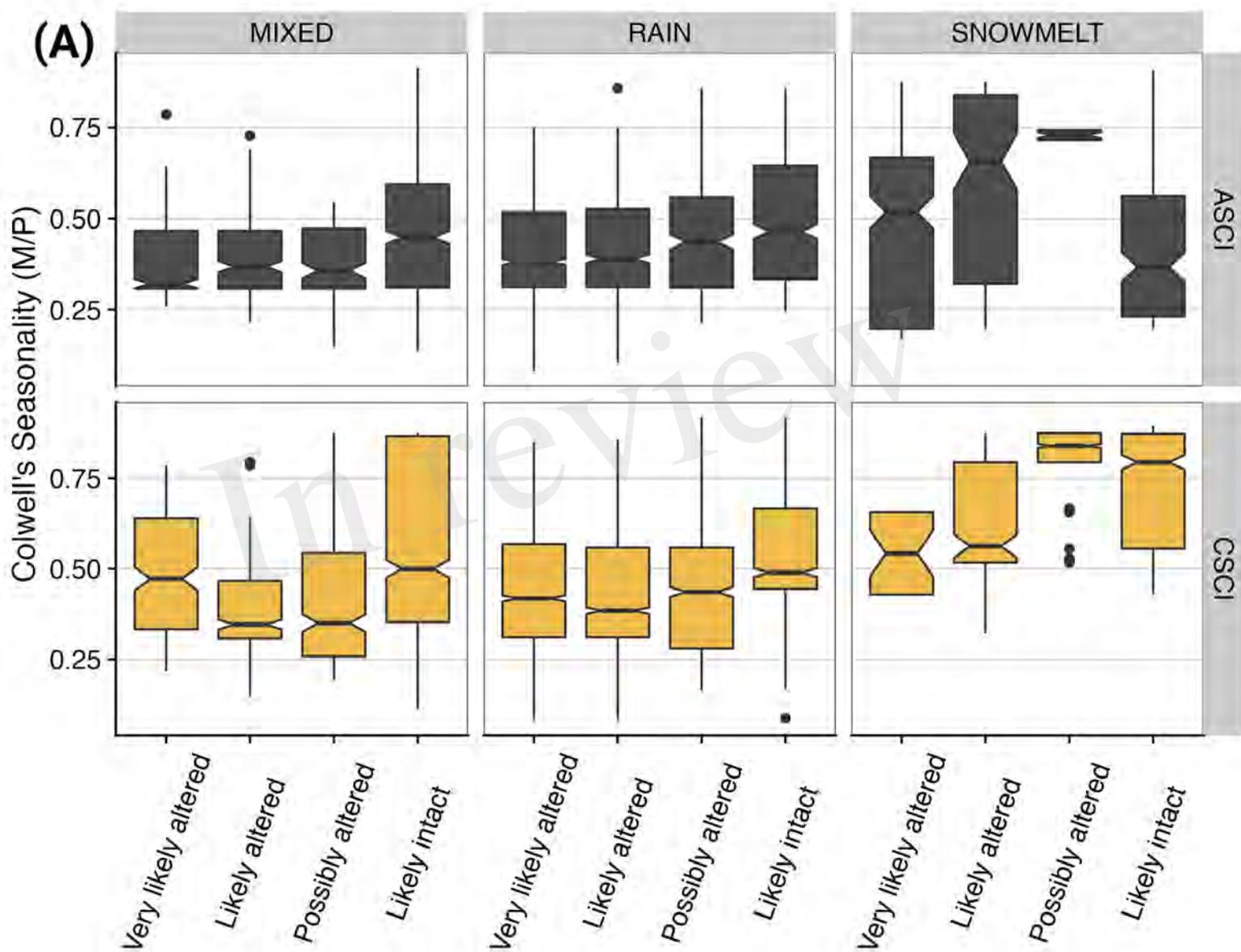


Figure 10.TIFF

